

1 Revisiting optimal foraging theory (OFT) in a changing Amazon: Implications for conservation  
2 and management

3 Brian M. Griffiths<sup>1\*</sup>, Mark Bowler<sup>2,3</sup>, Joseph Kolowski<sup>4</sup>, Jared Stabach<sup>5</sup>, Elizabeth L. Benson<sup>6</sup>, &  
4 Michael P. Gilmore<sup>1</sup>

5  
6 <sup>1</sup>School of Integrative Studies, George Mason University, 4400 University Drive, Fairfax, VA  
7 22030, USA

8 <sup>2</sup>School of Science, Technology and Engineering, University of Suffolk, Waterfront Building,  
9 Neptune Quay, Ipswich, IP4 1QJ, UK

10 <sup>3</sup>Suffolk Sustainability Institute, Waterfront Building, Neptune Quay, Ipswich, IP4 1QJ, UK

11 <sup>4</sup>Smithsonian Conservation Biology Institute, Smithsonian-Mason School of Conservation, 1500  
12 Remount Rd. Front Royal, VA 2263

13 <sup>5</sup>Smithsonian National Zoo & Conservation Biology Institute, 1500 Remount Rd, Front Royal,  
14 VA 22630

15 <sup>6</sup>OnePlanet, Inc., 2444 Navarre Way, Virginia Beach, VA 23456, USA

16 \*Corresponding author: Brian M. Griffiths, [bgriffi7@gmu.edu](mailto:bgriffi7@gmu.edu), 2 Heather Hill Ln., Elkton, MD,  
17 21921

18 Keywords: decision-making, game meat, mammal, Peru, prey selection, wild game

19 Abstract

20 Hunter decision-making influences prey selection and is key to understanding the impacts  
21 of hunting on biodiversity. Optimal foraging theory (OFT) is often used to describe the decision-  
22 making and prey selection of subsistence hunters. We examined the behavior and game meat use  
23 of hunters in an indigenous Amazonian community and used free-listing and generalized linear  
24 mixed-effects models under the framework of OFT to assess the decision-making of individuals  
25 who hunt for economic gain and subsistence. We found that prey selection generally followed  
26 OFT, and was influenced by hunter prowess, patch choice, and characteristics of the prey  
27 encountered. Hunters preferred paca (*Cuniculus paca*), collared peccary (*Pecari tajacu*), and  
28 brocket deer (*Mazama americana*) and only partially preferred tapir (*Tapirus terrestris*) and  
29 large-bodied primates likely due to economic influences such as access to markets and prices,  
30 contrary to OFT.

31 Keywords: decision-making, game meat, mammal, Peru, prey selection, wild game

32

### 33 Introduction

34           Tropical forests support an estimated 50% of the world's described species and many  
35 more that remain undescribed (Groombridge and Jenkins 2002). Logging and overhunting are  
36 frequently cited as drivers of biodiversity loss, particularly in the Amazon Basin of South  
37 America (Benítez-López et al. 2017; Brancalion et al. 2018; Milner-Gulland and Bennett 2003;  
38 Redford 1992; Schipper et al. 2008). Overhunting often results from commercial hunting, a  
39 consequence of local hunters selling game meat to local markets to feed demand in urban centers  
40 throughout the Amazon Basin (El Bizri *et al.* 2020; Mayor et al., 2021; Lozano and Fang 2004).  
41 While the effects of heavy hunting pressure are well studied (Benítez-López et al. 2017), the  
42 impacts of small-scale hunting on mammal populations are often confounded by other  
43 disturbances, such as habitat degradation (Peres 2001; Redford 1992, Remis and Jost Robinson  
44 2012). Even in the absence of other disturbances, variability can still be introduced due to  
45 differences in hunting practices and behavior of individual hunters and communities. Optimal  
46 foraging theory (OFT) (Charnov 1976) is often used to describe the decision-making behavior of  
47 hunter-gatherers. OFT is a set of models which dictate the breadth and proportion of prey items  
48 that a predator should take (diet breadth model) as well as where those prey should be taken  
49 (patch choice model) (Hames and Vickers 1982).

50           Under the diet breadth model (Emlen 1966), prey are ranked according to the return rate  
51 afforded to the hunter, often calculated as the caloric intake gained per hour spent pursuing,  
52 killing, and processing the animal (Alvard 1993; Bettinger et al. 2015; Smith *et al.* 1983;  
53 Winterhalder 1981). As such, OFT focuses mainly on hunting for subsistence, not economic  
54 gain. A prey item is included in the diet breadth if the return rate of pursuing and killing the  
55 animal is higher than the expected mean return rate of ignoring the species and continuing to

56 search for a higher ranked species, thereby maximizing the hunter's short-term harvesting rate  
57 (Alvard 1993; Bettinger et al. 2015; Hawkes and O'Connell 1992; Smith *et al.* 1983; Stephens  
58 and Krebs 1986). A key prediction of the diet breadth model is the zero-one rule, which states  
59 that hunters should always pursue species within the diet breadth and never pursue species  
60 outside, thereby never exhibiting "partial preferences" for prey (Bettinger et al. 2015; Levi et al.  
61 2011a; Stephens and Krebs 1986). However, studies have shown that partial preferences could  
62 be introduced by conservation behaviors, taboos, and economic influences (Alvard 1993; Hames  
63 2007). In the case where hunting causes declines of high-ranked species and the search time for  
64 those species increases, new, more abundant species may be added to the diet breadth (Bettinger  
65 et al. 2015; Hames and Vickers 1982; Pyke *et al.* 1977). Erosion of traditional taboos and  
66 technologies may also allow new species to be introduced to the diet breadth (Hames 2007;  
67 Hames and Vickers 1982; Hill and Hawkes 1983; Vickers 1980). Frequently, the species that  
68 provide the highest return rate are the largest bodied species, which are also vulnerable to  
69 extirpation due to low reproductive rates (Mayor et al. 2017; Redford 1992; Vickers 1991).

70 The patch choice model (Charnov 1976) assumes that hunters are central place foragers  
71 who exploit resources nearby the community first, eventually producing a gradient of game  
72 availability (Venkataraman et al. 2017; Winterhalder 2001). The model states that hunters will  
73 choose to hunt in patches where resource availability is highest, and therefore the return rate is  
74 highest (Bettinger et al. 2015). However, species which become rare in depleted patches are still  
75 pursued when encountered and are not dropped from the diet breadth (Bettinger et al. 2015;  
76 Hames and Vickers 1982; Winterhalder and Lu 1997). Hunters who do not change their patch  
77 choice in response to game depletion are forced to expand their diet breadth to include more  
78 abundant, less preferred species (Charnov 1976; Hames and Vickers 1982; Levi et al. 2011a).

79           The choices and characteristics of individual hunters or hunts may influence the  
80 effectiveness of OFT in predicting hunter behavior, including the hunter's skill (Hill *et al.* 1987),  
81 beliefs (Lemos *et al.* 2021), food preferences (Chaves *et al.* 2020) and cultural taboos (Hames  
82 2007), the conditions of the hunt (Levi *et al.* 2011), the characteristics of the animal encountered,  
83 and the available technology (Bettinger 2009; Hames and Vickers 1982). Individuals may also  
84 have different goals and currencies and therefore return rates (Mithen 1989). For example, a  
85 hunter may want meat for subsistence to feed his family, for gifting to neighbors to further his  
86 social standing (Bird and Bird 1997; Hawkes *et al.* 2001), or for sale to market for economic  
87 gain (Ayres *et al.* 1991). Previous studies on the applications of OFT to human hunters have  
88 aggregated data from individual hunters to be analyzed on a community level (e.g. Hurtado,  
89 Hawkes, and Kaplan 1985; Hill, Kaplan, Hawkes, and Hurtado 1985); however, the variation due  
90 to these individual influences is then lost (Asmhyr *et al.* 2013; Chaves *et al.* 2020; Mithen 1989).

91           This study revisits OFT by examining hunting behavior, partial preferences, patch choice,  
92 and game meat economics in an indigenous Amazonian community that has access to the  
93 regional market. We use free-listing of hunter preferences and generalized linear mixed-effects  
94 models to assess whether OFT applies and is accurate in describing individual hunter behavior  
95 where economics influences hunter decision-making alongside subsistence pressures.

96 Specifically, we address the following questions:

- 97           1. Which species are sold for economic gain by hunters and how much are they worth?
- 98           2. Which species do hunters prefer, and do those preferences align with what would be  
99           predicted by OFT?
- 100           3. Does OFT accurately predict individual hunter decision making upon encountering prey?

101

## 102 Methods

### 103 *Study Site*

104 Fieldwork was conducted in collaboration with the Maijuna (Orejón) indigenous group of  
105 the northeastern Peruvian Amazon. The Maijuna are a Western Tucanoan people with a  
106 population of approximately 600 individuals (Gilmore 2010). There are four Maijuna  
107 communities: Puerto Huamán and Nueva Vida along the Yanayacu River, Sucusari along the  
108 Sucusari River, and San Pablo de Totolla (Totoya) along the Algodón River (Fig. 1). These three  
109 river basins are part of the ancestral territory of the Maijuna and no other communities are  
110 located within this area (Gilmore 2010). The Maijuna traditionally lived in the interfluvial area  
111 between these three rivers. This residence pattern lasted until the early 1900s when the Maijuna  
112 began to slowly migrate downriver, due to influence from missionaries and *patrones* (colonists  
113 and their descendants who exploited indigenous labor to harvest forest resources), to where they  
114 eventually formed their current communities (Bellier 1993, 1994). Greater ethnographic context  
115 and a detailed ethnohistory of the Maijuna can be found in Gilmore (2010).

116 Fieldwork for this study was conducted in the Maijuna community of Sucusari  
117 (72.92995° W, 3.24373° S) (Fig. 1). Sucusari is approximately 126 km by river from the city of  
118 Iquitos, the commercial and political center of Loreto. However, the trip can be shortened to 70  
119 km by crossing the thin isthmus between the Napo and Amazon Rivers by road at Mazán, a small  
120 town. The titled land of the community encompasses 4,771 hectares and adjoins the Maijuna-  
121 Kichwa Regional Conservation Area (MKRCA), a 391,040-hectare protected area that is made  
122 up of Maijuna ancestral lands and collaboratively managed by Maijuna and Kichwa indigenous  
123 communities and the regional government (El Peruano, 2015). The community has a population  
124 of 166 residents made up of both monofamilial and plurifamilial households. Of all community

125 members, 59% are ethnically Maijuna, 35% are mestizos (individuals of mixed Amerindian and  
126 Iberian descent; Coomes and Ban 2004) and 6% are indigenous Kichwa (Roncal *et al.* 2018).

127         Subsistence and income generating strategies of community members include hunting,  
128 fishing, swidden-fallow agriculture, and the gathering of various non-timber forest products  
129 (Gilmore 2010). Community members sell game meat in the city of Iquitos, in towns  
130 surrounding their communities on the Napo River and in the market of Mazán (Gilmore 2020).  
131 Mean household income of Maijuna families averages about 505 USD annually (Horn *et al.*  
132 2012, Figure 2). Game meat is also gifted, a cultural norm where a hunter who has recently killed  
133 an animal may gift a portion to neighbors, family, and friends within the community. Only men  
134 hunt in Sucusari (Roncal *et al.* 2018), hunting opportunistically from canoes, on foot and at  
135 mineral licks (Gilmore *et al.* 2020). All hunters hunt with shotguns or with machetes for slow-  
136 moving game species. Two hunters in the community use dogs to hunt. Only community  
137 members from Sucusari had hunting rights in the Sucusari titled lands and MKRCA. Our  
138 research team has been working in the Sucusari community since 1999, which has allowed us to  
139 foster strong and close relationships with hunters in the community.

#### 140 *Data Collection and Analysis*

141         All aspects of this study were approved by George Mason University's Institutional  
142 Review Board, project #1288488-1. Prior informed consent (PIC) was obtained from the  
143 Sucusari community as well as from individual research participants before beginning the  
144 interviews for this study. We conducted weekly semi-structured interviews in Spanish (Berg and  
145 Lune 2014) with 19 hunters (90.48% of all active hunters) in Sucusari to capture decision-  
146 making processes on prey selection and the economics of hunting for a ten-month period from  
147 September 2018 to June 2019. The mean age of hunters interviewed was 41 years old, with a

148 range of 22 to 68. Of the 19 hunters interviewed, ten were ethnically Maijuna, eight were  
149 mestizo, and one was Kichwa. If a hunter was not home when we visited his house, we made at  
150 least two more attempts during the days immediately following the initial visit. If a hunter was  
151 still not at home after three visit attempts, we gathered the data for that week during the  
152 following week's interview. During the interview, for each hunt, we first asked hunters what  
153 time they went hunting, how long the hunt lasted for, and the mode(s) of travel they used (i.e.  
154 travel by boat, canoe, or on foot). We then asked which animals they encountered during each  
155 hunt and where, asking them to indicate the location on a basemap of locally relevant points, and  
156 if they had attempted to kill the animal. If they did not try to kill the animal, we asked why they  
157 chose not to. We digitized all encounter and kill locations, extracting coordinates using ArcGIS  
158 (ESRI, 2018). We coded interview responses focused on decision-making (Berg and Lune  
159 2014), where reasons for not shooting an animal were coded and then grouped into overall  
160 themes. These themes were "Killed", "Escaped", "Low Return", "Attack-Limited", and  
161 "Conservation", following classifications of behavior informed by Alvard (1993) and OFT  
162 (Table 1). We used these codes to calculate partial preferences based on pursuit rate, or the  
163 proportion of encounters in which the animal was killed, shot at, or pursued until it escaped.

164         When a hunter reported that he killed an animal, we asked how many kg of meat he sold,  
165 to whom he sold it, and at what price. Each hunter was also asked how many kg of meat he had  
166 consumed with his family and how many kg he had gifted and to whom. All hunters in the  
167 community have spring scales which they use to measure the mass of a carcass and portions to  
168 sell or gift.



169 We conducted separate semi-structured interviews in Spanish (Berg and Lune 2014) with  
170 17 of the 19 hunters at the end of the study period to provide context to observed pursuit rates  
171 and relative species-specific preferences. Two hunters were not interviewed because they moved  
172 away from the community for an extended period while the survey was conducted. We asked  
173 hunters to list three species they preferred to kill while hunting (a) by canoe, (b) by land, (c) at a  
174 mineral lick, and (d) for game meat to sell. These different hunting methods (a-c) were chosen  
175 because hunters noted that they look for different species using each technique. Freelisting is an  
176 interview method that can be used to determine the salience of named species (Quinlan 2005;  
177 Roncal *et al.* 2018). In this case, it allowed us to rank species in relation to individual hunter  
178 preferences. Following Quinlan (2005), we used the following formula to determine the salience  
179 of an individual species in a hunter's list of preferences:

$$180 \quad \text{Salience} = (1 + \text{length}_i - \text{position}_i) / \text{length}_i$$

181 where length is the total number of species listed by the hunter, and position is the  
182 numbered position at which species *i* appears in the hunter's list. We calculated the total salience  
183 of each species named at least three times for hunters in Sucusari as the mean of all salience  
184 values for that species. Species listed in free lists as preferred are considered those which hunters  
185 perceive as having the highest return rates, under OFT. Species which hunters did not list as  
186 preferred but pursued sometimes are considered species that are partially preferred (Alvard  
187 1993).

### 188 *Mixed-Effects Modeling*

189 We used generalized linear mixed-effects models to assess hunter decision-making using  
190 encounter data. Species that were encountered by hunters were aggregated into species groups

191 because of a lack of data for some species (Table S1). Paca (*Cuniculus paca*), collared peccary  
192 (*Pecari tajacu*), tapir (*Tapirus terrestris*), and brocket deer (*Mazama americana* and *Mazama*  
193 *gouazoubira*) formed their own separate groups, since these were species that were preferred by  
194 hunters. Game birds were grouped together, as were large-bodied (> 1.5 kg, those which are  
195 pursued by hunters) and small-bodied primates. All other species, which were hunted but not  
196 listed as preferred (such as the agouti (*Dasyprocta fuliginosa*) and kinkajou (*Potos flavus*), were  
197 grouped into “Other Species”. Encounters with carnivores (e.g. the jaguar (*Panthera onca*),  
198 puma (*Puma concolor*), and giant river otter (*Pteronura brasiliensis*)) were excluded from  
199 analyses because they were not killed during the study period. The yellow-footed tortoise  
200 (*Chelonoidis denticulata*) was also excluded from analyses because they were not killed with a  
201 shotgun.

202         As our primary, overarching model to evaluate hunter decision-making, we constructed a  
203 generalized linear mixed-effects model with a binomial distribution to assess the factors that  
204 contributed to whether a hunter decided to pursue an animal he encountered (Model 1). Each  
205 encounter was a sample ( $n = 1,012$ ), and the response variable was a 0/1 based on whether the  
206 hunter pursued the animal (animal was killed, hunter took a shot at the animal unsuccessfully, or  
207 hunter pursued the animal and it escaped). We included a series of covariates formed under the  
208 framework of OFT (Table 2) and the hunter’s name as a random effect to account for  
209 pseudoreplication caused by repeated samples from the same set of hunters. Only hunters who  
210 went hunting at least ten times during the study period were included in modeling ( $n = 17$ ).

211         To add context to our primary model’s results, we constructed three more models. The  
212 second model was created to evaluate the factors which contributed to a hunter deciding not to

213 pursue an animal because its return rate was perceived to be too low and was therefore outside  
214 the diet breadth (Model 2). This was a generalized linear mixed-effects models with a binomial  
215 distribution. Each encounter was considered a sample ( $n = 1,012$ ) and the response variable was  
216 a 0/1 based on whether the animal was ignored because of a reason coded as “low return”.  
217 Covariates included were the same as the previous model (Table 2), with the exception of species  
218 group which was excluded due to a lack of sufficient data in each group.

219         Since an assumption of OFT is that species are never partially pursued, we constructed a  
220 third generalized linear mixed-effects model with a binomial distribution to assess the factors  
221 that influence when hunters choose to pursue partially preferred species and test this assumption  
222 (species in the “Other Species”, “Small Primate”, “Large Primate” and “Game Birds” groups, see  
223 Table S1, which were not always pursued when encountered) (Model 3). Each encounter with a  
224 partially preferred species was a sample ( $n = 652$ ) and the response variable was a 0/1 based on  
225 whether the hunter pursued the animal (the animal was shot at, killed, or escaped before an  
226 intended shot was taken). Covariates included were the group size and mean body size of the  
227 animal, the Euclidean distance from the community, the total duration of the hunt, the hunter’s  
228 overall return rate, and the consumer/producer ratio of the hunter’s household, with the hunter’s  
229 name as a random effect to account for pseudoreplication (Table 2).

230         Under the hypothesis that a partial preference for a species could be a result of increased  
231 wariness of hunters close to the community, requiring greater effort on the part of the hunter to  
232 pursue species, we constructed a final generalized linear mixed-effects model with a binomial  
233 distribution to assess the factors which contributed to preferred, major game species (paca, tapir,  
234 collared peccary, and brocket deer) escaping hunters during an encounter (Model 4). Each

235 encounter with one of these species was a sample ( $n = 361$ ) and the response was a 0/1 based on  
236 whether or not the animal escaped while the hunter intended to kill it (i.e. the reason for not  
237 shooting the animal was that it escaped pursuit or a shot was taken unsuccessfully). Covariates in  
238 the model included the species group, the hunter's overall return rate, as a measure of the  
239 hunter's prowess, and the Euclidean distance from the community.

240 For all models, continuous covariates were scaled and checked for collinearity before  
241 including them in the model, with a correlation cutoff of 0.60 for inclusion (Dormann *et al.*  
242 2013) before proceeding with model selection. We used a model-averaging approach to  
243 determine the optimal model, forming a candidate set of all possible combinations of relevant  
244 covariates then weighting the coefficient estimates of each model by that model's Akaike weight  
245 (Burnham and Anderson 2002). All mixed-effects models were calculated using the *lme4* (Bates  
246 *et al.*, 2015) package and model averaging conducted using the MuMIn package (Barton 2009) in  
247 R (version 4.0.3) (R Core Team, 2020). Model fit was assessed using the theoretical pseudo- $R^2$   
248 for mixed-effects models for the top-ranked models in each averaged candidate set (Nakagawa *et*  
249 *al.* 2017) and by visually examining residuals.

250

## 251 Results

### 252 *General Hunting Behavior*

253 We collected data on 671 hunting trips during the study period. Overall, 38.6% of hunts  
254 ( $n = 256$ ) were conducted over land, 31.9% ( $n = 214$ ) by canoe, 27.0% ( $n = 181$ )  
255 opportunistically from motorboats, and 3.0% ( $n = 20$ ) included both a canoe and land

256 component. During 14.3% of hunts, a hunter visited a mineral lick. However, only 1.2% of these  
257 visits to a mineral lick lasted longer than 15 minutes. Overall, 66.9% of hunts ( $n = 449$ ) were  
258 conducted during the day, 20.3% of hunts ( $n = 136$ ) took place at night, and 12.8% of hunts ( $n =$   
259 86) spanned both night (20.00h – 06.00h) and daylight hours (06.00h – 20.00h). On average,  
260 hunts were 6.85 hours (median 6 hours) in length, with a range of ten minutes to 27 hours.

261 Of all game meat harvested during the study period, by weight, hunters sold 58.2%,  
262 gifted 10.6%, and kept 31.2%. The species that were killed most frequently were paca, collared  
263 peccary, and red brocket deer (Fig. 2). The white-lipped peccary (*Tayassu pecari*) was not  
264 encountered by any hunters during the study period because the local population had experienced  
265 a population crash (e.g., see [Fragoso 2004](#)) The species that hunters sold the highest proportion  
266 of meat from were paca, collared peccary, red brocket deer, and Brazilian tapir (Fig. 2).

267 Hunters had highly variable overall return rates, with a mean of 0.85 (SD = 0.54) kg meat  
268 harvested/hr spent hunting and a range of 0.19 to 1.94 kg meat harvested/hr spent hunting. All  
269 hunters sold meat during the study period and return rates from sales had a mean of 5.03 soles  
270 (1.51 USD) earned/hr spent hunting and a range of 0.51 to 16.77 soles (0.15 to 5.04 USD)  
271 earned/hr spent hunting (Table S2). Overall, the paca was the most frequently reported species  
272 with 212 encounters, followed by the agouti (103 encounters) and the kinkajou (98 encounters).  
273 Hunters reported encountering animals from 0.10 km from the community to 30.82 km from the  
274 community (mean = 9.20 km, SD = 7.73 km) (Fig S1). When broken down by species group,  
275 there was some variation in the distribution of encounters by distance from the community (Fig  
276 S2). Primates in particular, both large and small-bodied, were encountered more frequently

277 nearby the community whereas paca were generally found farther from the community compared  
278 to other groups.

### 279 *Game Meat Economics*

280 Game meat was frequently sold within the community to neighbors and friends, to  
281 communities nearby Sucusari on the Napo River (labeled as “Napo” in Fig. 3), and to the  
282 regional market in the town of Mazán or the city of Iquitos. Game meat was typically sold salted  
283 or salted and then smoked. Four species made up 63.2% of total meat sales by mass: paca,  
284 collared peccary, red brocket deer, and tapir. On average, meat from each of the most commonly  
285 sold species sold for the most money at the regional market, at S/. 12.12 (3.64 USD) for paca  
286 meat, S/. 11.94 (3.59 USD) for collared peccary, S/. 8.92 (2.68 USD) for red brocket deer, and  
287 S/. 9.80 (2.94 USD) for tapir per kg (Fig 3). Meat sold for less in nearby communities along the  
288 Napo River and went for even lower prices in the community itself (Fig. 3).

289 Hunters also frequently gifted game meat that they harvested to extended family, friends,  
290 and neighbors (10.6% of all meat harvested, by mass). The most commonly gifted species were  
291 the yellow-footed tortoise (36.36% by mass), grey brocket deer (29.55% by mass), South  
292 American coati (*Nasua nasua*) (18.60% by mass), and common woolly monkey (*Lagothrix*  
293 *lagotricha*) (17.65% by mass) (Fig. S3). All of these species were sold relatively infrequently  
294 (Fig. 2).

### 295 *Hunter Preferences*

296 During interviews at the end of the study period, we asked hunters to list three species  
297 they prefer to kill while hunting on trails, by canoe, and at mineral licks and three species they

298 prefer to kill when hunting for economic gain. The collared peccary and the paca were the most  
299 preferred species, with overall salience indices of 0.80 and 0.76, respectively (Table 3). The  
300 collared peccary was the most preferred species targeted on land with an index of 0.95 but was  
301 less preferred than other species by canoe or at mineral licks. The paca was the most preferred  
302 species targeted from canoes, with an index of 0.99, but was also less preferred than other  
303 species on land and at mineral licks. The white-lipped peccary and Brazilian tapir were the most  
304 preferred species targeted in mineral licks, with indices of 0.83 and 0.79 respectively, but were  
305 not preferred prey on land or in canoes. The collared peccary and paca were also the most highly  
306 preferred species when hunting for meat to sell, with salience indices of 0.88 and 0.72  
307 respectively while the white-lipped peccary and red brocket deer were similarly ranked at 0.42  
308 and 0.43 respectively. The tapir was never listed as a species that was targeted for sale. While the  
309 red brocket deer was listed repeatedly by hunters, it was the least preferred species overall  
310 compared to the other four species (Table 3).

### 311 *Partial Preferences of Species*

312 We used hunter-reported encounter data and the reasons the hunter gave for not shooting  
313 an animal during an encounter to assess partial preferences of species groups. The paca, collared  
314 peccary, and brocket deer were almost always pursued upon encounter, each with a pursuit rate  
315 of over 96.0% (Table 4). The tapir and game birds were usually pursued when encountered, at a  
316 rate of 76.2% and 73.5%, respectively. However, when the tapir was not pursued it was typically  
317 for an attack-limited reason (14.3% of encounters), while the reason game birds were not  
318 pursued was typically for a perceived low return rate (23.5% of encounters). The species in the  
319 “Other Species” category, which were species not listed as preferred species by hunters, were

320 killed at least once during the study period, and were not encountered frequently enough to form  
321 their own group, were also partially preferred at a pursuit rate of 52.5% of encounters (Table 4).  
322 Large primates were also partially preferred at a pursuit rate of 42.8% of encounters, but were  
323 often ignored for a perceived low return rate (33.1% of encounters) and for attack-limited (13.8%  
324 of encounters) and conservation reasons (10.3% of encounters). Small primates were almost  
325 never pursued (3.50% of encounters), typically for a perceived low return rate (65.9% of  
326 encounters) or for conservation reasons (22.0% of encounters) (Table 4).

### 327 *Mixed-Effects Modeling*

328 We used generalized linear mixed-effects models to assess the factors that influenced  
329 whether hunters decided to pursue an animal that they encountered while hunting (Model 1). The  
330 averaged model of hunter decision-making included species group and body size of the animal as  
331 significant covariates (Table 5). In general, predicted probabilities of pursuit followed stated  
332 pursuit rates (Table 4), with the paca, deer, and collared peccary almost always pursued (Fig. 4).  
333 Model results showed that as the hunter's overall return rate increased, the probability of pursuit  
334 decreased (Fig 4). Probability of pursuit increased with the body size of the animal and group  
335 size, and decreased with distance from the community and consumer/producer ratio (Table 5).

336 Model results for Model 2 showed that mean price was the most important factor in  
337 determining whether a hunter did not pursue a species because of a perceived low return rate  
338 (Table 5). As distance from the community increases, hunters are more likely to ignore species  
339 with cheaper meat prices, particularly for species worth less than about six Peruvian soles (1.81  
340 USD) per kg (Fig 5). Model results for Model 3 showed that species body size and mean price  
341 were the most important covariates (Table 5). Hunters were more likely to shoot at larger



342 partially preferred species and those that were worth more economically (Table 5). Model results  
343 for the probability of a preferred species group (paca, collared peccary, tapir, and deer) escaping  
344 during an encounter (Model 4) showed that both the hunter's overall return rate (a measure of  
345 hunter prowess) and the distance from the community were important factors (Table 5). As  
346 distance from the community and the hunter's overall return rate increased, the probability of an  
347 animal escaping decreased (Fig 6, Table 5). There was no correlation between the hunter's  
348 overall return rate and the distance from the community (0.048). Parameters of model subsets for  
349 all averaged models (Models 1-4) are given in Table S3.

350

## 351 Discussion

### 352 *Diet Breadth Model*

353 Results from hunter preference data indicated clear preferences for some species over  
354 others. Interestingly, the white-lipped peccary and tapir had lower preference indices in  
355 comparison to the paca and collared peccary, even though these species are reported to be among  
356 the most commonly hunted species in Amazonia (de Andrade Melo *et al.* 2015; El Bizri *et al.*  
357 2020; Mayor *et al.* 2021). These results match game meat prices, where paca and collared  
358 peccary are the species most likely to be sold and the most valuable. Accordingly, the paca and  
359 collared peccary were the most commonly sold species by proportion of total meat sales. Some  
360 hunters remarked that the tapir was often not killed because of its large body size, with monetary  
361 rewards not matching the effort required to process and carry the meat to market. These results  
362 indicate that hunters consider both economic and subsistence return rates when they are hunting,  
363 challenging the theory that the largest species provide the highest return rates to hunters (Alvard

364 1993; Bettinger *et al.* 2015). The relatively low salience for the paca from land-based hunting  
365 can likely be partially attributed to the nocturnal nature of the species (Griffiths *et al.* 2020) or  
366 differential habitat preferences (El Bizri *et al.* 2018). Several hunters noted in their interviews  
367 that they did not like to hunt by land at night because they were afraid of being bitten by snakes.  
368 It's also possible that the paca is more commonly found in riparian areas at night, leading hunters  
369 to target them by canoe rather than by land.

370 Hunter preferences found here did not line up with previously reported results from other  
371 communities, such as those reported by Bodmer (1995), where tapirs and white-lipped peccaries  
372 were the most preferred species. It's possible that since the white-lipped peccary was not  
373 encountered by hunters during the study period, hunters' listed preferences were conflated,  
374 placing the white-lipped peccary lower on the preference list than it would be if it were still  
375 abundant.

376 Evaluation of hunter pursuit rates for species groups revealed that hunters in Sucusari do  
377 not follow the predicted zero-one rule of the diet breadth model (Bettinger *et al.* 2015; Stephens  
378 and Krebs 1986), similar to what was reported by Alvard (1993) where Piro hunters frequently  
379 exhibited only partial preferences for some species. The measured pursuit rates we present  
380 showed that paca, collared peccary, and brocket deer were in the optimal diet breadth of all  
381 hunters, and were almost always pursued, a result that is directly in line with free-listed  
382 preferences. Since these species were almost always pursued, it's likely that these species yield  
383 the highest return rate for hunters. The currency of the return rate (Winterhalder 1981) in this  
384 case is likely a mixture of economic gain and subsistence, where the most preferred species are  
385 not only relatively large-bodied but also economically valuable.

386           The tapir was only a partially preferred species, pursued only on 76.2% of encounters,  
387 even though it is the largest game species. Given the reasons for not pursuing tapirs on some  
388 occasions (9.5% low return, 14.3% attack-limited), it's possible that hunters are not willing to  
389 invest the time needed to transport and process the tapir, a significant task given its size, since  
390 that handling time precludes time that could be spent searching for other smaller and more  
391 economically profitable species (Bettinger *et al.* 2015; Chaves *et al.*, 2020; Smith *et al.* 1983;  
392 Winterhalder and Lu 1997). It also should be noted that tapir often could not be effectively killed  
393 with a shotgun without getting close or using a special lead slug that some hunters carry. In some  
394 cases, hunters may have encountered tapir without having a slug with them, increasing the  
395 probability of attack limitation (Levi *et al.* 2011). Tapir also had a high preference index (0.79)  
396 to be hunted at mineral licks, but only 14.31% of hunts during the study period included a visit to  
397 a mineral lick. This result lends evidence to the idea that while hunters know where to find tapir,  
398 which frequently visit mineral licks in the region (Griffiths *et al.* 2020), they choose not to hunt  
399 them in favor of other species. Large-bodied primates were also partially preferred and were  
400 frequently ignored because of a perceived low return rate (33.1% of encounters) or for  
401 conservation reasons (10.3% of encounters). The prevalence of conservation-based reasons for  
402 ignoring primates while hunting contrasts with results presented by (Alvard 1993, 1995), which  
403 showed that Piro hunters targeted species only with short-term maximization of harvest in mind  
404 and not long-term conservation. Levi *et al.* (2009; 2011b) estimated a kill rate for another large-  
405 bodied primate, the spider monkey (*Ateles chamek*) at 0.90, much larger than the 0.119 we report  
406 here. Tapir and large primates perform key ecosystem functions, including seed dispersal (Brodie  
407 *et al.* 2009; Effiom *et al.* 2014; Galetti *et al.* 2001; Tobler 2008) and are crucial to the health of  
408 the broader ecosystem. The partial preference results indicate that hunters were not going to the

409 forest to look for these species specifically, which may result in lower offtake levels and a lower  
410 risk of overexploitation of those species in our study area than others across the region (Peres  
411 1990) if the population of the community stays low (Alvard *et al.* 1997). In contrast, the  
412 relatively high pursuit rate and preference for the paca compared to the tapir has conservation  
413 implications since the paca is a rodent with a much higher reproductive rate than that of the tapir  
414 and is therefore more resistant to population decline (Bodmer *et al.* 1997; El Bizri *et al.* 2018).  
415 However, our data did show fewer encounters with paca nearby the community, suggesting that  
416 some local defaunation or behavioral avoidance may be occurring.

417 Our decision-making model and partial preference results showed that species were  
418 pursued based on hunter and species-specific characteristics, showing that individual hunters  
419 have different diet breadths that are influenced by their own beliefs (e.g. propensity to ignore  
420 primates for conservation reasons) (Lemos *et al.* 2021), and their own skill at hunting (Hill *et al.*  
421 1987). These results support arguments made by Hames and Vickers (1982) and Mithen (1989),  
422 who stated that individuals vary in their behavior and return rates and therefore their diet  
423 breadths, and studies assessing hunter behavior should examine individual choices rather than  
424 aggregating data for analysis at the group level.

#### 425 *Patch Choice Model*

426 The importance of distance from the community in all decision-making models indicated  
427 that hunters are making decisions about potential return rates based upon the patch that they  
428 enter, consistent with the patch choice model (Charnov 1976) and results presented by Hames  
429 and Vickers (1982), where hunters will choose to enter patches further from the community for  
430 preferred species or will accept a wider variety of species closer to the community. Levi *et al.*

431 (2011a) also showed an effect of distance from the community on diet breadth, where hunters  
432 expanded their diet breadth as distance increased to avoid the risk of returning empty-handed.  
433 Our model results indicate preferred species were more likely to escape encounters closer to the  
434 community, indicating that those animals are more wary of hunters closer to the community and  
435 exhibit a behavioral response to humans that enables them to escape more often. The probability  
436 of escape was much lower farther from the community and when facing a more skilled hunter.  
437 Since there was no correlation between hunter skill and distance from the community of  
438 encounter locations, it can be assumed that this result is not due to more skilled hunters choosing  
439 patches farther from the community. The wariness of species closer to the community likely has  
440 a large effect on hunter return rate, since hunters would need to spend more time pursuing  
441 preferred species when they are encountered close to the community. This result supports the  
442 patch choice theory, where hunters who are hunting closer to the community likely have to  
443 accept a wider range of species than those hunting farther away (Hames and Vickers 1982).  
444 However, the patch choice theory has focused on depletion of game following the central place  
445 foraging theory causing a decline in return rate in patches closer to the community, not a  
446 behavioral response by animals (Charnov 1976; Pyke *et al.* 1977; Winterhalder 2001). Our  
447 results for increased wariness nearby the community add greater context to the expansion of diet  
448 breadth predicted by the patch choice model.

#### 449 *Conclusions*

450 We suggest that the behavior of Amazonian hunters who have access to markets but also  
451 still depend on hunting for subsistence is still largely explained by the optimal foraging theory  
452 and its applications, the diet breadth model and the patch choice model. With greater access to

453 technology and markets, it's likely that other Amazonian communities which previously hunted  
454 only for subsistence now hunt for both subsistence and economic gain, and that these results of  
455 prey selection are generalizable to those communities. The results of this study indicate that the  
456 influence of the market may drive preference towards valuable species such as the paca and  
457 collared peccary, and away from tapir and large primates which have been shown to be well  
458 within the diet breadth of subsistence hunters in other communities (Alvard 1993). The partial  
459 preference for these keystone (Paine 1995) seed dispersers compared to species such as the paca  
460 suggests that indigenous communities engaged in small-scale commercial and subsistence  
461 hunting may have a lesser effect on biodiversity than has been previously suggested by other  
462 studies showing tapir and primates as preferred species (Alvard 1993; Hames and Vickers 1982;  
463 Ojasti 1984). These results also suggest that future fluctuations in market prices could alter these  
464 preferences and further endanger these species or continue to further reduce their preference by  
465 hunters.

466

#### 467 Acknowledgements

468 We would like to acknowledge the Maijuna community of Sucusari for their support and  
469 collaboration on this research. We would like to thank OnePlanet, Inc. and Explorama Lodges  
470 who provided in-kind support for the project. We would like to thank the Fulbright Association  
471 for providing a Fulbright U.S. Student Grant to BMG to complete this research.

472

#### 473 Compliance with Ethics Statement

474 All aspects of this study were approved by George Mason University's Institutional  
475 Review Board, project #1288488-1. Prior informed consent (PIC) was obtained from the  
476 community as well as from individual research participants before beginning this study.

477

#### 478 Conflict of Interest Statement

479 The authors have no conflicts of interest to declare.

480

#### 481 Funding

482 This research was funded by a Fulbright U.S. Student Grant awarded to BMG.

483

#### 484 Data Availability Statement

485 The datasets generated during and/or analyzed during the current study are not publicly  
486 available since specific hunters are identifiable by the information in interviews but are available  
487 from the corresponding author on reasonable request.

488

489 References

- 490 Alvard, M., 1995. Shotguns and sustainable hunting in the Neotropics. *Oryx* 29, 58–66.  
491 doi:<https://doi.org/10.1017/S0030605300020883>
- 492 Alvard, M.S., 1993. Testing the “ecologically noble savage” hypothesis: Interspecific prey  
493 choice by Piro hunters of Amazonian Peru. *Human Ecology* 21, 355–387.
- 494 Alvard, M., Robinson, J.G., Redford, K.H., Kaplan, H., 1997. The sustainability of subsistence  
495 hunting in the neotropics. *Conserv. Biol.* 11, 977–982. doi:[https://doi.org/10.1046/j.1523-](https://doi.org/10.1046/j.1523-1739.1997.96047.x)  
496 [1739.1997.96047.x](https://doi.org/10.1046/j.1523-1739.1997.96047.x)
- 497 Asmyhr, L., Willebrand, T., Hörnell-Willebrand, M., 2013. The optimal foraging theory,  
498 crowding and Swedish grouse hunters. *Eur J Wildl Res* 59, 743–748.  
499 doi:<https://doi.org/10.1007/s10344-013-0729-4>
- 500 Ayres, J.M., Lima, D. de M., Martins, E. de S., Barreiros, J.L.K., 1991. On the track of the road:  
501 changes in subsistence hunting in a Brazilian Amazonian village. *Neotropical wildlife use*  
502 *and conservation* 82, 92.
- 503 Barton, K., 2009. Mu-MIn: Multi-model inference. R Package Version 0.12.2/r18. [http://R-](http://R-Forge.R-project.org/projects/mumin/)  
504 [Forge.R-project.org/projects/mumin/](http://R-Forge.R-project.org/projects/mumin/)
- 505 Bates D., Mächler M., Bolker B., Walker S., 2015. Fitting linear mixed-effects models using  
506 lme4. *Journal of Statistical Software* 67, 1–48.  
507 doi: [10.18637/jss.v067.i01](https://doi.org/10.18637/jss.v067.i01).
- 508 Benítez-López, A., Alkemade, R., Schipper, A.M., Ingram, D.J., Verweij, P.A., Eikelboom,  
509 J.A.J., Huijbregts, M.A.J., 2017. The impact of hunting on tropical mammal and bird



- 510 populations. *Science* (New York, N.Y.) 356, 180–183.
- 511 doi:<https://doi.org/10.1126/science.aaj1891>
- 512 Berg, B.L., Lune, H., 2014. *Qualitative research methods for the social sciences*. Harlow.
- 513 Bettinger, R. L., 2009. *Hunter-gatherer foraging: Five simple models*. Clinton Corners: Eliot
- 514 Werner Publications.
- 515 Bettinger, R.L., Garvey, R., Tushingham, S., 2015. Hunter-gatherers as optimal foragers, in:
- 516 Bettinger, R.L., Garvey, R., Tushingham, S. (Eds.), *Hunter-Gatherers: Archaeological*
- 517 *and Evolutionary Theory, Interdisciplinary Contributions to Archaeology*. Boston, MA,
- 518 pp. 91–138. doi:[https://doi.org/10.1007/978-1-4899-7581-2\\_4](https://doi.org/10.1007/978-1-4899-7581-2_4)
- 519 Bird, R.L.B., Bird, D.W., 1997. Delayed reciprocity and tolerated theft: The behavioral ecology
- 520 of food-sharing strategies. *Current anthropology* 38, 49–78.
- 521 Bodmer, R.E., 1995. Managing Amazonian wildlife: Biological correlates of game choice by
- 522 detribalized hunters. *Ecological Applications* 5, 872–877.
- 523 doi:<https://doi.org/10.2307/2269338>
- 524 Bodmer, R.E., Eisenberg, J.F., Redford, K.H., 1997. Hunting and the likelihood of extinction of
- 525 Amazonian mammals. *Conservation Biology* 11, 460–466.
- 526 doi:<https://doi.org/10.1046/j.1523-1739.1997.96022.x>
- 527 Brancalion, P.H.S., Almeida, D.R.A. de, Vidal, E., Molin, P.G., Sontag, V.E., Souza, S.E.X.F.,
- 528 Schulze, M.D., 2018. Fake legal logging in the Brazilian Amazon. *Science Advances* 4,
- 529 eaat1192. doi:<https://doi.org/10.1126/sciadv.aat1192>
- 530 Brodie, J.F., Helmy, O.E., Brockelman, W.Y., Maron, J.L., 2009. Bushmeat poaching reduces
- 531 the seed dispersal and population growth rate of a mammal-dispersed tree. *Ecological*
- 532 *Applications* 19, 854–863.

- 533 Burnham, K.P., Anderson, D.R., 2002. Model selection and multimodel inference: A practical  
534 information-theoretic approach, 2nd ed. New York. doi:<https://doi.org/10.1007/b97636>
- 535 Charnov, E.L., 1976. Optimal foraging, the marginal value theorem. *Theoretical population*  
536 *biology* 9, 129–136.
- 537 Chaves, L.S., Alves, R.R.N., Albuquerque, U.P., 2020. Hunters' preferences and perceptions as  
538 hunting predictors in a semiarid ecosystem. *Science of The Total Environment* 726,  
539 138494. doi:<https://doi.org/10.1016/j.scitotenv.2020.138494>
- 540 Coomes, O. T., & Ban, N. (2004). Cultivated plant species diversity in home gardens of an  
541 Amazonian peasant village in northeastern Peru. *Economic Botany*, 58(3), 420–434.  
542 [https://doi.org/10.1663/0013-0001\(2004\)058\[0420:CPSDIH\]2.0.CO;2](https://doi.org/10.1663/0013-0001(2004)058[0420:CPSDIH]2.0.CO;2)
- 543 de Andrade Melo, E.R., Gadelha, J.R., Domingos da Silva, M. de N., da Silva Junior, A.P.,  
544 Mendes Pontes, A.R., 2015. Diversity, abundance and the impact of hunting on large  
545 mammals in two contrasting forest sites in northern amazon. *Wildlife Biol.* 21, 234–245.  
546 doi:<https://doi.org/10.2981/wlb.00095>
- 547 Dormann, C.F., Elith, J., Bacher, S., Buchmann, C., Carl, G., Carré, G., Marquéz, J.R.G.,  
548 Gruber, B., Lafourcade, B., Leitao, P.J., 2013. Collinearity: a review of methods to deal  
549 with it and a simulation study evaluating their performance. *Ecography* 36, 27–46.
- 550 Effiom, E.O., Birkhofer, K., Smith, H.G., Olsson, O., 2014. Changes of community composition  
551 at multiple trophic levels due to hunting in Nigerian tropical forests. *Ecography* 37, 367–  
552 377. doi:<https://doi.org/10.1111/j.1600-0587.2013.00359.x>
- 553 El Bizri, H.R., Fa, J.E., Bowler, M., Valsecchi, J., Bodmer, R., Mayor, P., 2018. Breeding  
554 seasonality in the lowland paca (*Cuniculus paca*) in Amazonia: interactions with rainfall,

- 555           fruiting, and sustainable hunting. *J Mammal* 99, 1101–1111.
- 556           doi:<https://doi.org/10.1093/jmammal/gyy102>
- 557 El Bizri, H.R., Morcatty, T.Q., Valsecchi, J., Mayor, P., Ribeiro, J.E., Vasconcelos Neto, C.F.,  
558           Oliveira, J.S., Furtado, K.M., Ferreira, U.C., Miranda, C.F., 2020. Urban wild meat  
559           consumption and trade in central Amazonia. *Conservation Biology* 34, 438–448.
- 560 El Peruano, 2015. Establecen área de conservación regional Maijuna Kichwa, ubicada en el  
561           departamento de Loreto. Decreto No. 008-2015. [http://](http://www.elperuano.com.pe/NormasElperuano/2015/06/17/1252025-2.html)  
562           [www.elperuano.com.pe/NormasElperuano/2015/06/17/1252025-2.html](http://www.elperuano.com.pe/NormasElperuano/2015/06/17/1252025-2.html).
- 563 Emlen, J.M., 1966. The role of time and energy in food preference. *The American Naturalist*  
564           100, 611–617.
- 565 Environmental Systems Research Institute (ESRI), 2018. ArcGIS release 10.6. Redlands, CA.
- 566 Fragoso, J.M., 2004. A long-term study of white-lipped peccary (*Tayassu pecari*) population  
567           fluctuation in northern Amazonia. *People in nature, wildlife conservation in South and*  
568           *Central America* 286–296.
- 569 Galetti, M., Keuroghlian, A., Hanada, L., Morato, M.I., 2001. Frugivory and seed dispersal by  
570           the lowland tapir (*Tapirus terrestris*) in southeast Brazil. *Biotropica* 33, 723–726.
- 571 Gilmore, M.P., 2010. The Maijuna: past, present, and future. Perú: Maijuna, Rapid biological  
572           and social inventories Report 22, 226.
- 573 Gilmore, M.P., Griffiths, B.M., Bowler, M., 2020. The socio-cultural significance of mineral  
574           licks to the Maijuna of the Peruvian Amazon: implications for the sustainable  
575           management of hunting. *Journal of Ethnobiology and Ethnomedicine* 16, 1–10.

- 576 Griffiths, B.M., Bowler, M., Gilmore, M.P., Luther, D., 2020. Temporal patterns of visitation of  
577 birds and mammals at mineral licks in the Peruvian Amazon. *Ecology and evolution* 10,  
578 14152–14164.
- 579 Groombridge, B., Jenkins, M., 2002. *World atlas of biodiversity: earth's living resources in the*  
580 *21st century*. Berkeley, Calif. : London.
- 581 Hames, R., 2007. Game conservation or efficient hunting?, in: *Evolutionary Perspectives on*  
582 *Environmental Problems*.
- 583 Hames, R.B., Vickers, W.T., 1982. Optimal diet breadth theory as a model to explain variability  
584 in Amazonian hunting. *American Ethnologist* 9, 358–378.
- 585 Hawkes, K., O'Connell, J., 1992. On optimal foraging models and subsistence transitions.  
586 *Current Anthropology* 33, 63–66. doi:<https://doi.org/10.1086/204035>
- 587 Hawkes, K., O'Connell, J.F., Jones, N.B., 2001. Hadza meat sharing. *Evolution and Human*  
588 *Behavior* 22, 113–142.
- 589 Hill, K., Hawkes, K., 1983. Neotropical hunting among the Ache of eastern Paraguay, in:  
590 *Adaptive responses of native Amazonians*. New York, pp. 139–188.
- 591 Hill, K., Kaplan, H., Hawkes, K., Hurtado, A.M., 1987. Foraging decisions among Ache hunter-  
592 gatherers: new data and implications for optimal foraging models. *Ethology and*  
593 *Sociobiology* 8, 1–36.
- 594 Hill, K., Kaplan, H., Hawkes, K., and Hurtado, A. (1985). Mens' time allocation to subsistence  
595 activities among the Ache of Eastern Paraguay. *Human Ecology* 13: 29-47.
- 596 Horn, C.M., Gilmore, M.P., Endress, B.A., 2012. Ecological and socio-economic factors  
597 influencing aguaje (*Mauritia flexuosa*) resource management in two indigenous  
598 communities in the Peruvian Amazon. *Forest Ecology and Management* 267, 93–103.

- 599 Hurtado, A., Hawkes, K., Hill, K., and Kaplan, H. (1985). Female subsistence strategies among  
600 the Ache hunter gatherers of Eastern Paraguay. *Human Ecology* 13: 1-28.
- 601 Lemos, L.P., Loureiro, L.F., Morcatty, T.Q., Fa, J.E., de Vasconcelos Neto, C.F.A., de Souza  
602 Jesus, A., da Silva, V.C., de Oliveira Ramalho, M.L., de Matos Mendes, A., Valsecchi, J.,  
603 El Bizri, H.R., 2021. Social correlates of and reasons for primate meat consumption in  
604 central Amazonia. *Int J Primatol* 42, 499–521. doi:[https://doi.org/10.1007/s10764-021-](https://doi.org/10.1007/s10764-021-00214-6)  
605 [00214-6](https://doi.org/10.1007/s10764-021-00214-6)
- 606 Levi, T., Lu, F., Yu, D.W., Mangel, M., 2011a. The behaviour and diet breadth of central-place  
607 foragers: an application to human hunters and Neotropical game management. *Evol Ecol*  
608 *Res* 13, 171–185.
- 609 Levi, T., Shepard, G.H., Ohl-Schacherer, J., Peres, C.A., Yu, D.W., 2009. Modelling the long-  
610 term sustainability of Indigenous hunting in Manu National Park, Peru: Landscape-scale  
611 management implications for Amazonia. *Journal of Applied Ecology* 46, 804–814.
- 612 Levi, T., Shepard, G.H., Ohl-Schacherer, J., Wilmers, C.C., Peres, C.A., Yu, D.W., 2011b.  
613 Spatial tools for modeling the sustainability of subsistence hunting in tropical forests.  
614 *Ecological Applications* 21, 1802–1818.
- 615 Lozano, E.P., Fang, T.G., 2004. Economic analysis of wildlife use in the Peruvian Amazon, in:  
616 *People in nature, wildlife conservation in South and Central America*. pp. 191–208.
- 617 Mayor, P., Bizri, H.R.E., Morcatty, T.Q., Moya, K., Bendayán, N., Solis, S., Neto, C.F.A.V.,  
618 Kirkland, M., Arevalo, O., Fang, T.G., Pérez-Peña, P.E., Bodmer, R.E., n.d. Wild meat  
619 trade over the last 45 years in the Peruvian Amazon. *Conservation Biology* n/a.  
620 doi:<https://doi.org/10.1111/cobi.13801>

- 621 Mayor, P., El Bizri, H., Bodmer, R.E., Bowler, M., 2017. Assessment of mammal reproduction  
622 for hunting sustainability through community-based sampling of species in the wild.  
623 Conservation Biology 31, 912–923.
- 624 Milner-Gulland, E.J., Bennett, E.L., 2003. Wild meat: the bigger picture. Trends in Ecology &  
625 Evolution 18, 351–357. doi:[https://doi.org/10.1016/S0169-5347\(03\)00123-X](https://doi.org/10.1016/S0169-5347(03)00123-X)
- 626 Mithen, S.J., 1989. Modeling hunter-gatherer decision making: Complementing optimal foraging  
627 theory. Human Ecology 17, 59–83.
- 628 Nakagawa, S., Johnson, P.C., Schielzeth, H., 2017. The coefficient of determination  $R^2$  and  
629 intra-class correlation coefficient from generalized linear mixed-effects models revisited  
630 and expanded. Journal of the Royal Society Interface 14, 20170213.
- 631 Ojasti, J., 1984. Hunting and conservation of mammals in Latin America. Acta Zoologica  
632 Fennica[ACTA ZOOL. FENN.]. 1984.
- 633 Paine, R.T., 1995. A conversation on refining the concept of keystone species. Conservation  
634 biology 9, 962–964.
- 635 Peres, C.A., 1990. Effects of hunting on western Amazonian primate communities. Biological  
636 Conservation 54, 47–59. doi:[https://doi.org/10.1016/0006-3207\(90\)90041-M](https://doi.org/10.1016/0006-3207(90)90041-M)
- 637 Peres, C.A., 2001. Synergistic effects of subsistence hunting and habitat fragmentation on  
638 Amazonian forest vertebrates. Conserv. Biol. 15, 1490–1505.  
639 doi:<https://doi.org/10.1046/j.1523-1739.2001.01089.x>
- 640 Pyke, G.H., Pulliam, H.R., Charnov, E.L., 1977. Optimal foraging: a selective review of theory  
641 and tests. The quarterly review of biology 52, 137–154.
- 642 Quinlan, M., 2005. Considerations for collecting freelists in the field: Examples from  
643 ethobotany. Field Methods 17, 219–234. doi:<https://doi.org/10.1177/1525822X05277460>

- 644 R Core Team, 2019. R: A language and environment for statistical computing. R Foundation for  
645 Statistical Computing, Vienna, Austria. URL [https://www.R- project.org/](https://www.R-project.org/).
- 646 Redford, K.H., 1992. The empty forest. *BioScience* 42, 412–422.  
647 doi:<https://doi.org/10.2307/1311860>
- 648 Remis, M.J., Jost Robinson, C.A., 2012. Reductions in primate abundance and diversity in a  
649 multiuse protected area: Synergistic impacts of hunting and logging in a Congo basin  
650 forest. *Am. J. Primatol.* 74, 602–612. doi:<https://doi.org/10.1002/ajp.22012>
- 651
- 652 Roncal, C.M., Bowler, M., Gilmore, M.P., 2018. The ethnoprimateology of the Maijuna of the  
653 Peruvian Amazon and implications for primate conservation. *Journal of ethnobiology and*  
654 *ethnomedicine* 14, 19.
- 655 Schipper, J., Chanson, J.S., Chiozza, F., Cox, N.A., Hoffmann, M., Katariya, V., Lamoreux, J.,  
656 Rodrigues, A.S.L., Stuart, S.N., Temple, H.J., Baillie, J., Boitani, L., Lacher, T.E.,  
657 Mittermeier, R.A., Smith, A.T., Absolon, D., Aguiar, J.M., Amori, G., Bakkour, N.,  
658 Baldi, R., Berridge, R.J., Bielby, J., Black, P.A., Blanc, J.J., Brooks, T.M., Burton, J.A.,  
659 Butynski, T.M., Catullo, G., Chapman, R., Cokeliss, Z., Collen, B., Conroy, J., Cooke,  
660 J.G., Fonseca, G.A.B. da, Derocher, A.E., Dublin, H.T., Duckworth, J.W., Emmons, L.,  
661 Emslie, R.H., Festa-Bianchet, M., Foster, M., Foster, S., Garshelis, D.L., Gates, C.,  
662 Gimenez-Dixon, M., Gonzalez, S., Gonzalez-Maya, J.F., Good, T.C., Hammerson, G.,  
663 Hammond, P.S., Happold, D., Happold, M., Hare, J., Harris, R.B., Hawkins, C.E.,  
664 Haywood, M., Heaney, L.R., Hedges, S., Helgen, K.M., Hilton-Taylor, C., Hussain, S.A.,  
665 Ishii, N., Jefferson, T.A., Jenkins, R.K.B., Johnston, C.H., Keith, M., Kingdon, J., Knox,

666 D.H., Kovacs, K.M., Langhammer, P., Leus, K., Lewison, R., Lichtenstein, G., Lowry,  
667 L.F., Macavoy, Z., Mace, G.M., Mallon, D.P., Masi, M., McKnight, M.W., Medellín,  
668 R.A., Medici, P., Mills, G., Moehlman, P.D., Molur, S., Mora, A., Nowell, K., Oates,  
669 J.F., Olech, W., Oliver, W.R.L., Oprea, M., Patterson, B.D., Perrin, W.F., Polidoro, B.A.,  
670 Pollock, C., Powel, A., Protas, Y., Racey, P., Ragle, J., Ramani, P., Rathbun, G., Reeves,  
671 R.R., Reilly, S.B., Reynolds, J.E., Rondinini, C., Rosell-Ambal, R.G., Rulli, M., Rylands,  
672 A.B., Savini, S., Schank, C.J., Sechrest, W., Self-Sullivan, C., Shoemaker, A., Sillero-  
673 Zubiri, C., De Silva, N., Smith, D.E., Srinivasulu, C., Stephenson, P.J., van Strien, N.,  
674 Talukdar, B.K., Taylor, B.L., Timmins, R., Tirira, D.G., Tognelli, M.F., Tsytsulina, K.,  
675 Veiga, L.M., Vié, J.-C., Williamson, E.A., Wyatt, S.A., Xie, Y., Young, B.E., 2008. The  
676 status of the world's land and marine mammals: Diversity, threat, and knowledge.  
677 *Science* 322, 225–230.

678 Smith, E.A., Bettinger, R.L., Bishop, C.A., Blundell, V., Cashdan, E., Casimir, M.J.,  
679 Christenson, A.L., Cox, B., Dyson-Hudson, R., Hayden, B., 1983. Anthropological  
680 applications of optimal foraging theory: a critical review [and comments and reply].  
681 *Current Anthropology* 24, 625–651.

682 Stephens, D.W., Krebs, J.R., 1986. *Foraging theory*.

683 Tobler, M.W., 2008. The ecology of the lowland tapir in Madre de Dios, Peru: using new  
684 technologies to study large rainforest mammals.

685 Venkataraman, V.V., Kraft, T.S., Dominy, N.J., Endicott, K.M., 2017. Hunter-gatherer  
686 residential mobility and the marginal value of rainforest patches. *PNAS* 114, 3097–3102.  
687 doi:<https://doi.org/10.1073/pnas.1617542114>



- 688 Vickers, W.T., 1991. Hunting yields and game composition over ten years in an Amazon Indian  
689 territory. *Neotropical wildlife use and conservation* 400, 53–81.
- 690 Vickers, W.T., 1980. An analysis of Amazonian hunting yields as a function of settlement age.  
691 *Working papers on south American Indians* 2, 7–29.
- 692 Winterhalder, B., 2001. The behavioural ecology of hunter gatherers.
- 693 Winterhalder, B., 1981. Optimal foraging strategies and hunter-gatherer research in  
694 anthropology: Theory and models.
- 695 Winterhalder, B., Lu, F., 1997. A forager-resource population ecology model and implications  
696 for indigenous conservation. *Conservation Biology* 11, 1354–1364.
- 697
- 698

699 Figure and Table Captions

700 Table 1. Themes of reasons hunters gave for not shooting at an animal during an encounter,  
701 grouped according to Alvard (1993) and the optimal foraging theory, in an indigenous  
702 community in the Peruvian Amazon.

703

704 Table 2. Hypotheses for inclusion of covariates in generalized linear mixed-effects models,  
705 formed under the framework of the optimal foraging theory, a description of how metrics were  
706 calculated, and mean and standard deviation of continuous covariates.

707

708 Table 3. Salience indices by species from semi-structured interviews with hunters in Sucusari,  
709 Peru, showing preferred species. Only species listed more than ten times by hunters are shown.

710

711 Table 4. Proportions of encounters that resulted in pursuit or not shooting at the animal because  
712 of a perceived low return rate, attack-limited reasons, or conservation reasons by species group.  
713 Species that were killed, were shot at unsuccessfully, or escaped are classified as pursued.

714

715 Figure 1. Study area, including the Maijuna community of Sucusari and the Maijuna-Kichwa  
716 Regional Conservation Area (MKRCA).

717

718 Figure 2. Proportions of total meat by mass that was sold, number of animals killed, and number  
719 of animals encountered broken down by species and ordered by number of kills. Only species  
720 that were killed more than five times during the study period are shown.

721 Note: \* indicates a preferred game species, as determined by salience indices.

722

723 Figure 3. Average price per kilogram of meat sold of the four most frequently sold species for  
724 sales within the community of Sucusari, Peru, to other communities nearby on the Napo River,  
725 and to the regional markets in either Mazán or Iquitos. Only species for which more than 30  
726 kilograms were sold during the study period are shown. Species are shown in descending order  
727 of frequency of kills. Descriptive error bars show ranges.

728

729 Figure 4. Generalized linear mixed-effects model results predicting the probability of a hunter  
730 pursuing an animal that is encountered versus the hunter's overall return rate and species group  
731 of the encountered animal. Other covariates held at the mean value for display, and hunter  
732 chosen for display represents the median intercept of the random effects of the optimal model.

733

734 Figure 5. Generalized linear mixed-effects model results of the predicted probability of a hunter  
735 not pursuing an animal upon encounter versus distance from the community and the mean price  
736 per kg of the meat of the species encountered. Hunter chosen for display represents the median  
737 intercept of the random effects of the optimal model. Mean price values chosen for display  
738 represent quartiles and mean of mean price.

739

740 Figure 6. Generalized linear mixed-effects model results of the predicted probability of a  
741 preferred species escaping a hunter during an encounter versus distance from the community and  
742 the hunter's overall return rate (a measure of hunter prowess). Hunter chosen for display  
743 represents the median intercept of the random effects of the optimal model.

*This version of the article has been accepted for publication, after peer review (when applicable) and is subject to Springer Nature's [AM terms of use](#), but is not the Version of Record and does not reflect post-acceptance improvements, or any corrections. The Version of Record is available online at: <https://doi.org/10.1007/s10745-022-00320-w>*

745 Table 1.  
746

Code	Explanation	Sample Interview Responses
Killed	The animal was killed without the use of a shotgun	I chased it into its hole and killed it with my machete
Escaped	The hunter intended to kill the animal and may have pursued it, but the animal escaped	It ran away and left me behind It escaped from me
Low Return	Killing the animal would have been too much effort for the return it gave	It was too small It was too hard to get to
Attack-Limited	The hunter is restricted in the number of kills he can make, and a shot would have precluded a later kill	I did not want to make any noise I did not have enough cartridges
Conservation	The hunter gave up a short-term gain to avoid killing certain species/sexes/age groups	It had young with it I don't eat monkeys

747

748

749 Table 2.

Fixed Effects	Hypothesis for Model 1	Reference	Calculation	Mean	SD
Distance from Community (km)	Hunters may seek patches farther from the community for a greater return rate (patch choice theory)	Charnov 1976; Alvard 1994	Euclidean distance from encounter to community center	9.13	7.70
Hunt Duration (hr)	When a hunt lasts longer unsuccessfully, hunters may be more apt to shoot less preferred species so they do not come home empty handed	Alvard 1993; Levi et al. 2011	Total duration of hunt in which hunter was actively searching for or pursuing prey	6.85	4.48
Group Size (# individuals)	Hunters may be more likely to shoot at larger groups, when the chances of success are higher	Alvard 1993	Estimated group size of animals	4.65	9.13
Hunter's Overall Return Rate (kg/hr)	Variation in hunter skill may drive variation in diet breadth	Hill et al. 1987	Total kg harvested over entire study period / total hours spent actively hunting	0.85	0.54
Hunter's Household Consumer/Producer Ratio	Hunters with larger households may have a wider diet breadth since food is needed more often	Alvard 1993	Total number of people in hunter's household / total number of hunters in the household	0.29	0.12
Body Size (kg)	Larger species may be more likely to be in the diet breadth of hunters because they provide a greater return rate	Alvard 1993	Estimated body size of species, from the literature, in kg	10.59	23.26
Species Group	Hunters may prefer some species over others due to factors such as taste	Alvard 1993	-	-	-
Distance from Community : Mean Price	Hunters may seek patches farther from the community specifically to target species that are more economically valuable (patch choice theory)	Charnov 1976; Alvard 1994	Mean price calculated as the mean number of Peruvian soles received per kg for all sales within that species group during the study period	-	-
Hunter's Overall Return Rate : Mean Price	Hunters that are more successful may be more likely to sell game meat, and less skilled hunters may be more willing to accept less valuable species	-	-	-	-

750  
751

752

753

754 Table 3.

Species	Hunting Method				Overall Mean
	Land	Canoe	Mineral Lick	Sale	
<i>P. tajacu</i>	0.95	0.89	0.48	0.88	0.80
<i>C. paca</i>	0.56	0.99	0.77	0.72	0.76
<i>T. terrestris</i>	0.53	0.50	0.79	-	0.60
<i>T. pecari</i>	0.63	0.25	0.83	0.42	0.53
<i>M. americana</i>	0.62	0.49	0.56	0.43	0.52

755

756

757

758 Table 4.

Species Group	Number of Encounters	Proportion Pursued*	Proportion Killed	Proportion Shot At Unsuccessfully	"Escaped"	Proportion of Encounters: Reasons for No Pursuit		
						"Low Return"	"Attack-Limited"	"Conservation"
Paca	212	0.972	0.726	0.042	0.203	0.009	0.019	0.000
Collared Peccary	78	0.962	0.526	0.038	0.397	0.000	0.038	0.000
Deer	50	0.960	0.540	0.040	0.380	0.020	0.020	0.000
Tapir	21	0.762	0.381	0.286	0.095	0.095	0.143	0.000
Game Birds	34	0.735	0.353	0.088	0.294	0.235	0.029	0.000
Other Species	304	0.526	0.234	0.056	0.237	0.280	0.069	0.125
Large Primate	143	0.420	0.119	0.105	0.196	0.336	0.140	0.105
Small Primate	171	0.035	0.012	0.012	0.012	0.661	0.088	0.216

759 \*Proportion pursued is the sum of proportions killed, shot at unsuccessfully, and “escaped”



Table 5. Generalized linear mixed-effects model results of hunter decision-making. Coefficient estimates of averaged models shown with standard error in parentheses. NA indicates that fixed effect was not tested in the model due to convergence issues, and a dash indicates that fixed effect was not tested because of the hypotheses of the model. Reference species group for species group fixed effect is the collared peccary. Statistically significant estimates shown in bold.

	<b>Model 1</b>	<b>Model 2</b>	<b>Model 3</b>	<b>Model 4</b>
<b>Response Variable (0/1)</b>	Whether an animal is pursued	Whether an animal is not pursued for perceived low return rate	Whether a partially preferred species is pursued	Whether an animal escapes pursuit by a hunter
<b>Sample of encounters</b>	All	All	Partially preferred species	Preferred species
<b>Sample Size (n)</b>	1,012	1,012	651	361
<b>Fixed Effects</b>				
Group Size	0.022 (0.074)	-0.005 (0.043)	-0.142 (0.134)	-
Body Size	<b>1.948 (0.775)</b>	-0.071 (0.166)	<b>3.062 (1.036)</b>	-
Distance from Community	-0.073 (0.113)	-0.067 (0.122)	-0.164 (0.198)	<b>-0.635 (0.146)</b>
Hunt Duration	-0.122 (0.131)	0.003 (0.056)	-0.166 (0.153)	-
Hunter's Overall Return Rate	-0.144 (0.188)	-0.021 (0.128)	-0.283 (0.267)	<b>-1.019 (0.260)</b>
Hunter's Consumer/Producer Ratio	-0.021 (0.109)	0.005 (0.111)	0.024 (0.124)	-
Species Group (Paca)	<b>2.057 (0.909)</b>	NA	NA	-0.256 (0.367)
Species Group (Deer)	-0.134 (0.945)	NA	NA	0.110 (0.326)
Species Group (Tapir)	<b>13.100 (4.563)</b>	NA	NA	0.111 (0.410)
Species Group (Game Birds)	-0.331 (1.064)	NA	NA	-
Species Group (Other Species)	-1.403 (0.915)	NA	NA	-
Species Group (Large Primate)	<b>-2.082 (0.955)</b>	NA	NA	-
Species Group (Small Primate)	<b>-4.750 (1.100)</b>	NA	NA	-
Mean Price	NA	<b>-1.631 (0.142)</b>	<b>1.002 (0.246)</b>	-
Mean Price:Distance from Community	NA	-0.203 (0.180)	-0.171 (0.236)	-
Mean Price:Hunter's Overall Return Rate	NA	-0.012 (0.056)	-0.170 (0.218)	-
<b>Random Effect</b>	Hunter Name	Hunter Name	Hunter Name	Hunter Name

<b>Marginal R<sup>2*</sup></b>	0.630	0.444	0.248	0.278
<b>Conditional R<sup>2*</sup></b>	0.670	0.534	0.360	0.397

\*R<sup>2</sup> values calculated based on the top-ranked model of each averaged model, by Akaike weight



Figure 1.

*This version of the article has been accepted for publication, after peer review (when applicable) and is subject to Springer Nature's [AM terms of use](#), but is not the Version of Record and does not reflect post-acceptance improvements, or any corrections. The Version of Record is available online at: <https://doi.org/10.1007/s10745-022-00320-w>*

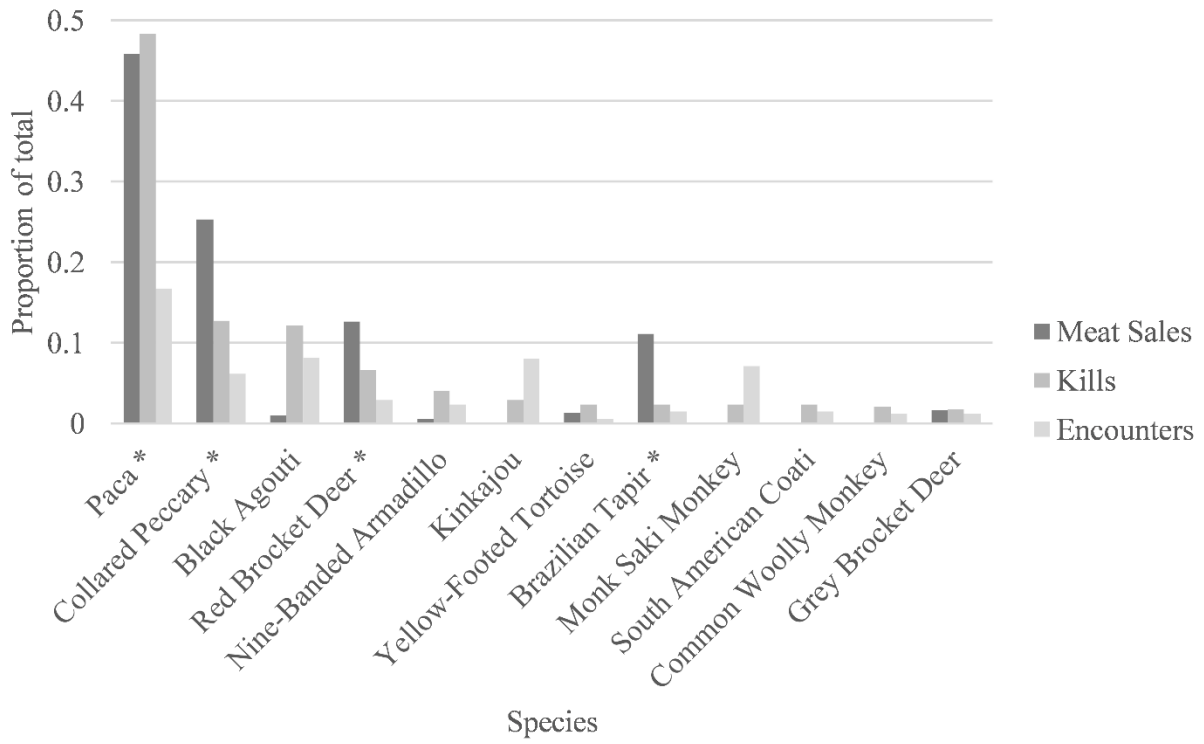


Figure 2.

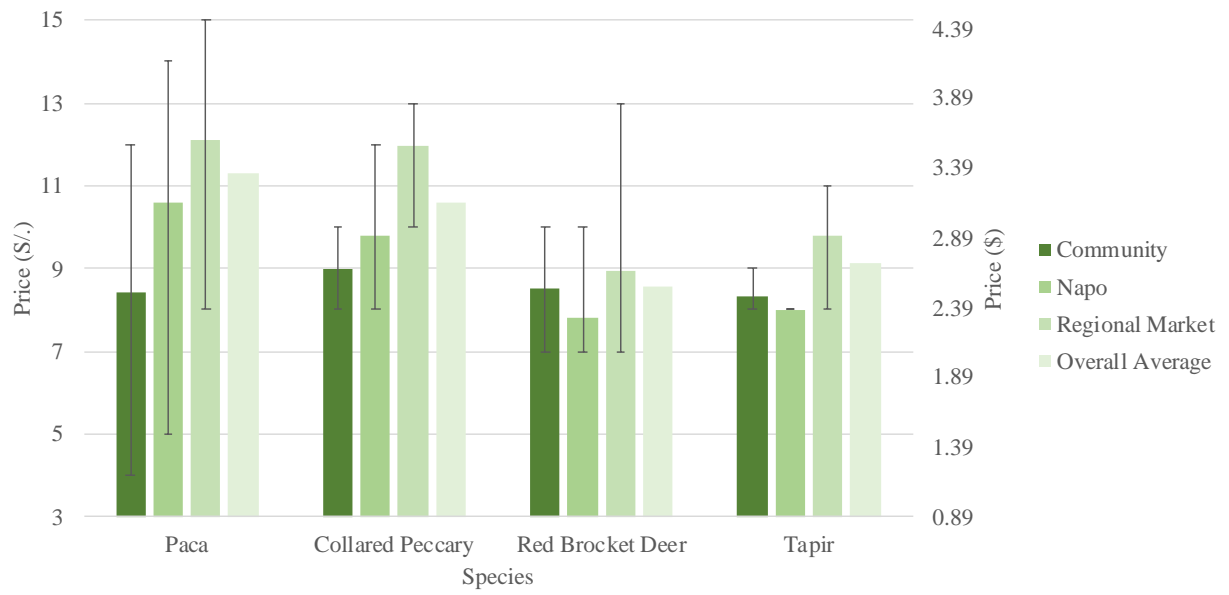


Figure 3.

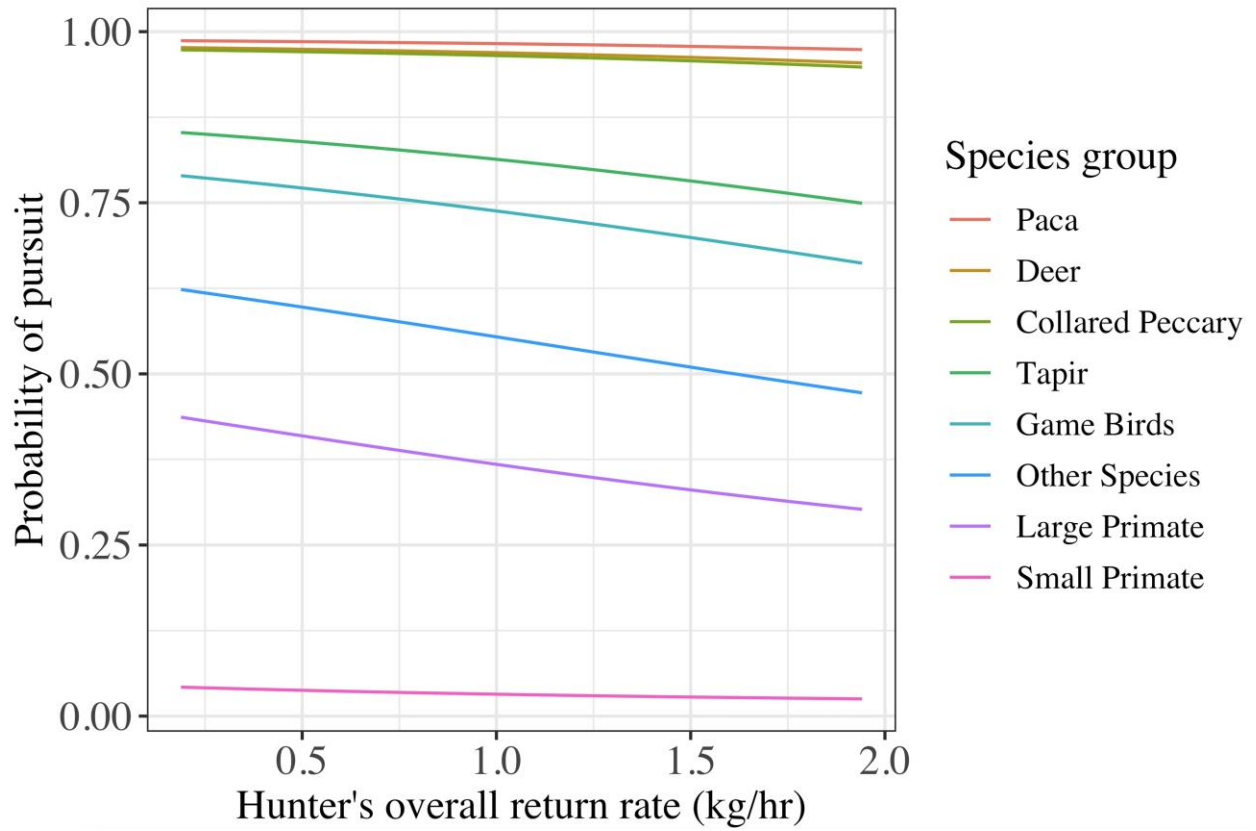


Figure 4.

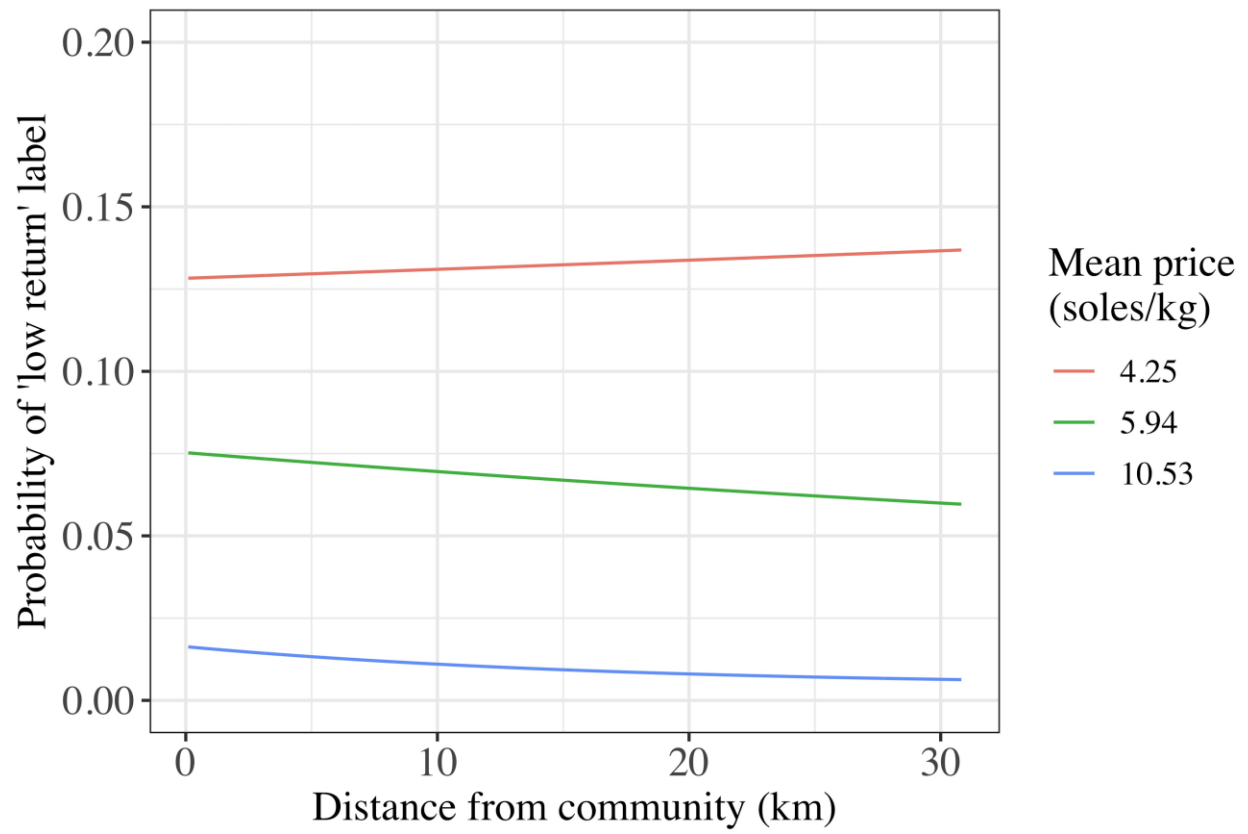


Figure 5.



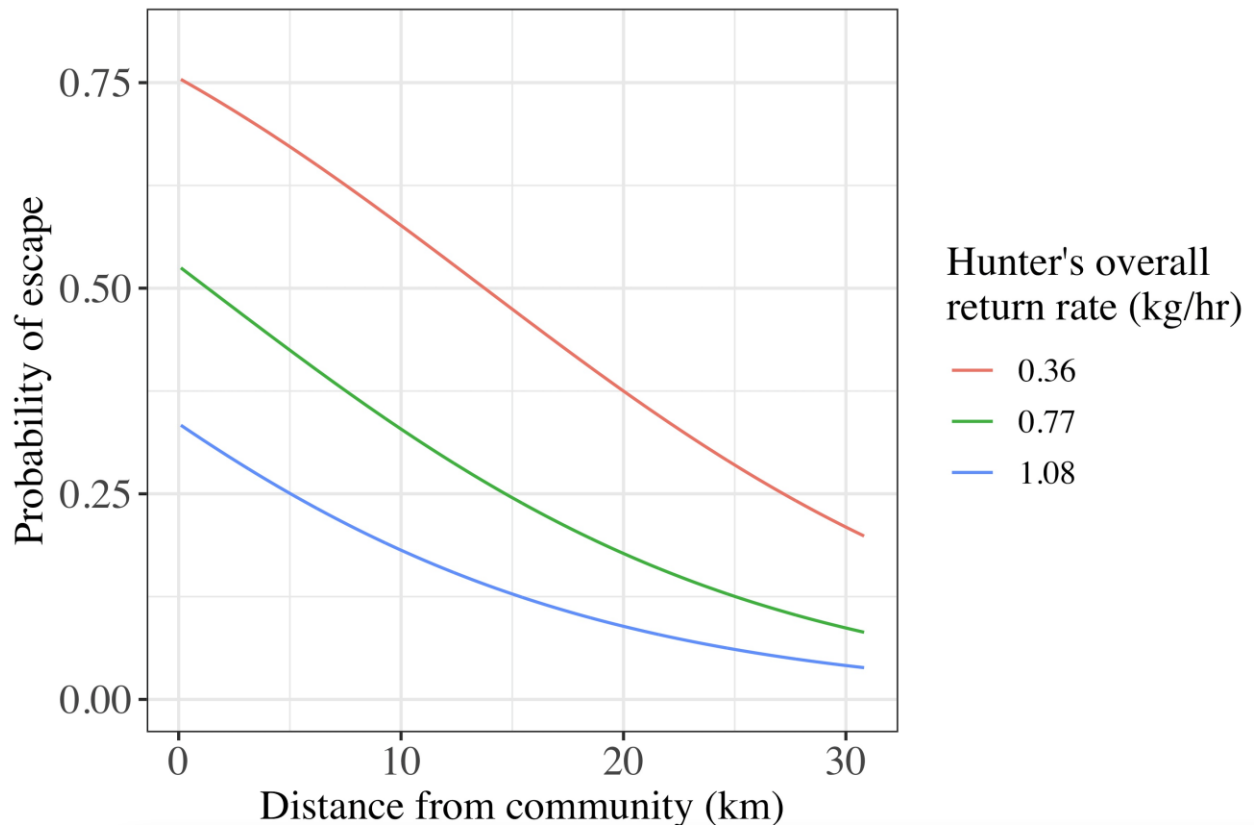


Figure 6.

## Supplementary Information

Table S1. Species included in generalized linear mixed-effects models, along with the species group they were placed in, the value for mean body size and mean price used for the species/species group, and the reference from which mean body size values were taken.

Class	Scientific Name	Common Name	Species Group	Mass (kg)	Mean Price (Soles/kg)	Reference
Mammal	<i>Pecari tajacu</i>	Collared Peccary	Collared Peccary	25.00	10.58	Bodmer et al., 2004
Mammal	<i>Mazama americana</i>	Red Brocket Deer	Deer	33.00	7.96	Bodmer et al., 2004
Mammal	<i>Mazama gouazoubira</i>	Grey Brocket Deer Common Woolly	Deer	17.35	7.96	Robinson & Redford, 1986
Mammal	<i>Lagothrix lagotricha</i>	Monkey	Large Primate	11.00	4.25	Bodmer et al., 2004
Mammal	<i>Alouatta seniculus</i>	Red Howler Monkey	Large Primate	6.19	4.25	Robinson & Redford, 1986
Mammal	<i>Cebus albifrons</i>	White-fronted Capuchin	Large Primate	2.01	4.25	Robinson & Redford, 1986
Mammal	<i>Pithecia monachus</i>	Monk Saki Monkey	Large Primate	1.80	4.25	Robinson & Redford, 1986
Mammal	<i>Hydrochoerus hydrochaeris</i>	Capybara	Other Species	30.00	5.25	Bodmer et al., 2004
Mammal	<i>Myrmecophaga tridactyla</i>	Giant Anteater	Other Species	27.00	5.25	Robinson & Redford, 1986
Mammal	<i>Dasyprocta fuliginosa</i>	Black Agouti	Other Species	5.00	5.25	Bodmer et al., 2004
Mammal	<i>Tamandua tetradactyla</i>	Southern Tamandua Linnaeus's Two-toed	Other Species	4.56	5.25	Robinson & Redford, 1986
Mammal	<i>Choloepus didactylus</i>	sloth	Other Species	4.15	5.25	Robinson & Redford, 1986
Mammal	<i>Nasua nasua</i>	South American Coati	Other Species	3.88	5.25	Robinson & Redford, 1986
Mammal	<i>Dasytus novemcinctus</i>	Nine-banded Armadillo	Other Species	3.54	5.25	Robinson & Redford, 1986
Mammal	<i>Potos flavus</i>	Kinkajou	Other Species	2.49	5.25	Robinson & Redford, 1986
Mammal	<i>Cuniculus paca</i>	Paca	Paca	9.00	11.32	Bodmer et al., 2004
Mammal	<i>Callicebus moloch</i>	Dusky Titi Monkey Yellow-handed Titi	Small Primate	1.17	0	Robinson & Redford, 1986
Mammal	<i>Callicebus lucifer</i>	Monkey	Small Primate	1.00	0	Robinson & Redford, 1986
Mammal	<i>Aotus vociferans</i>	Spix's Night Monkey	Small Primate	0.87	0	Robinson & Redford, 1986

Mammal	<i>Saimiri sciureus</i>	Common Squirrel Monkey	Small Primate	0.69	0	Robinson & Redford, 1986
Mammal	<i>Leontocebus nigricollis</i>	Black-mantled Tamarin	Small Primate	0.36	0	Robinson & Redford, 1986
Mammal	<i>Tapirus terrestris</i>	Brazilian Tapir	Tapir	160.00	9.11	Bodmer et al., 2004
Bird	<i>Mitu salvini</i>	Salvin's Curassow Blue-throated Piping Guan	Game Birds	3.06	0	Begazo & Bodmer, 1998
Bird	<i>Pipile cumanensis</i>	Spix's Guan	Game Birds	1.30	0	Begazo & Bodmer, 1998
Bird	<i>Penelope jacquacu</i>	Nocturnal Curassow	Game Birds	1.28	0	Begazo & Bodmer, 1998
Bird	<i>Nothocrax urumtum</i>	Tinamou	Game Birds	1.25	0	Kattan et al. 2016
Bird	<i>Tinamidae sp.</i>	Grey-winged Trumpeter	Game Birds	1.10	0	Dunning, 2007
Bird	<i>Psophia crepitans</i>	Speckled Chachalaca	Game Birds	1.03	0	Dunning, 2007
Bird	<i>Ortalis guttata</i>		Game Birds	0.50	0	Begazo & Bodmer, 1998

Table S2. Return rates of 17 hunters in Sucusari, Peru, measured by different metrics based on the outcomes of the meat each hunter harvested during the study period. Return rate of total kg/hr was used as the covariate for “hunter return rate” in generalized linear mixed-effects models.

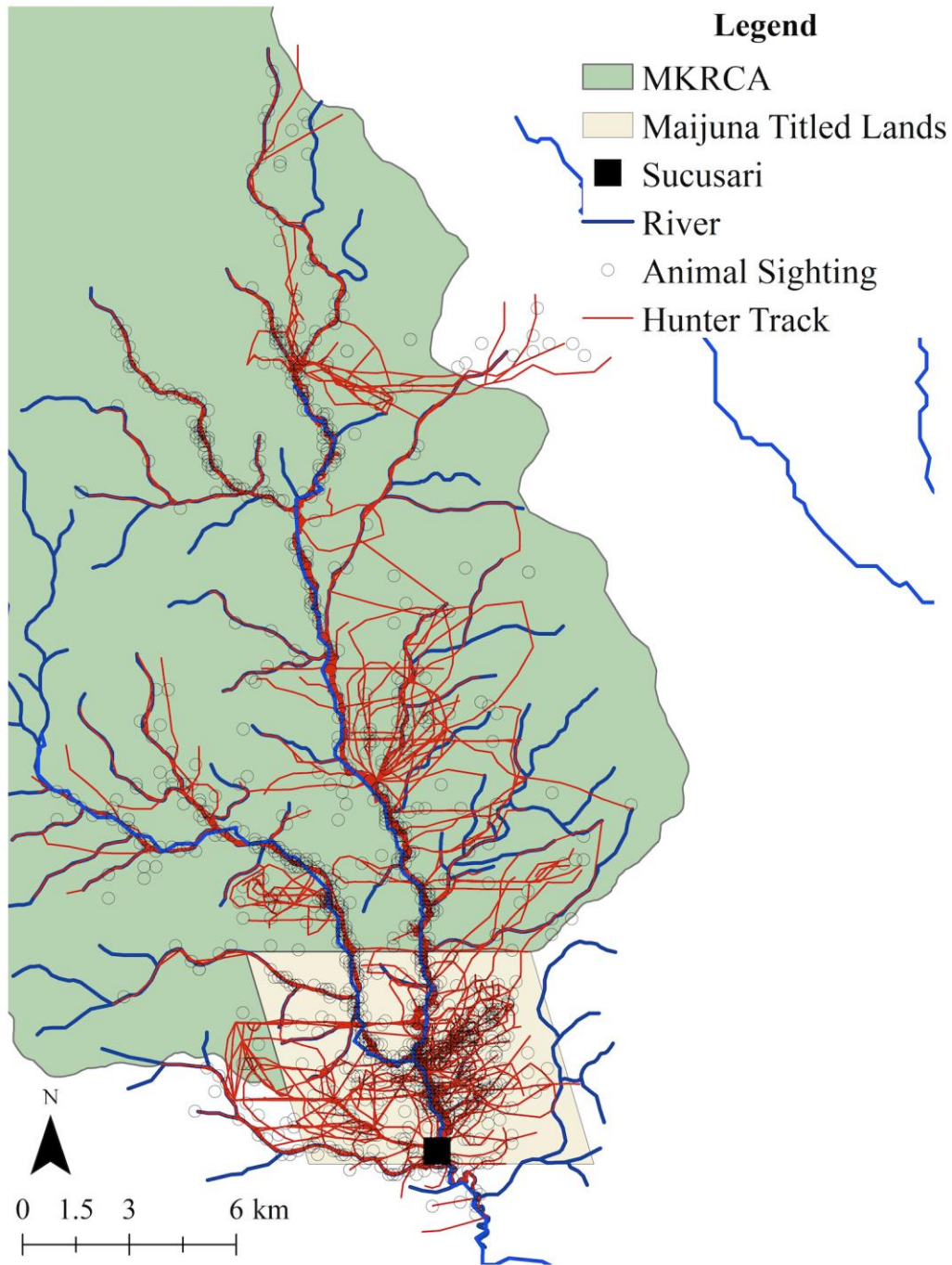
Hunter	Total kg/hr	Soles Earned/hr	kg Sold/hr	kg Gifted/hr	kg Kept for Consumption/hr
1	0.188	0.511	0.060	0.049	0.079
2	0.196	0.664	0.061	0.070	0.065
3	0.319	1.570	0.171	0.005	0.142
4	0.324	1.317	0.131	0.086	0.107
5	0.360	1.696	0.173	0.027	0.160
6	0.408	3.929	0.350	0.032	0.025
7	0.505	2.322	0.200	0.037	0.269
8	0.680	1.593	0.202	0.030	0.449
9	0.856	5.910	0.528	0.056	0.272
10	0.882	1.658	0.153	0.284	0.444
11	1.042	9.636	0.910	0.056	0.075
12	1.080	0.966	0.094	0.055	0.931
13	1.195	2.648	0.257	0.421	0.517
14	1.252	9.323	1.007	0.120	0.126
15	1.520	16.769	1.483	0.000	0.036
16	1.616	10.947	1.181	0.006	0.429
17	1.941	14.118	1.412	0.000	0.529
Mean	0.845	5.034	0.493	0.079	0.274
SD	0.536	5.161	0.498	0.111	0.245

Table S3. Generalized linear mixed-effects model results of hunter decision-making showing the covariates of the top-ranked models of each model series, number of parameters (K),  $\Delta$ AIC, and Akaike weight ( $w$ ) used in model averaging. All candidate models within 2 AIC points of the top-ranked model are shown.

Model	K	$\Delta$ AIC	$w$
<i>Series 1</i>			
Body Size + Hunt Duration + Species Group + Success Rate	12	0.00	0.095
Body Size + Hunt Duration + Species Group + Success Rate + Distance from Community	13	0.49	0.074
Body Size + Hunt Duration + Species Group	11	0.50	0.074
Body Size + Hunt Duration + Species Group + Distance from Community	12	0.96	0.059
Body Size + Species Group	10	1.20	0.052
Body Size + Species Group + Distance from Community	11	1.29	0.050
Body Size + Species Group + Success Rate	11	1.56	0.043
Body Size + Species Group + Success Rate + Distance from Community	12	1.62	0.042
Body Size + Species Group + Success Rate + Group Size	13	1.71	0.040
Body Size + Species Group + Success Rate + Consumer/Producer Ratio	13	1.83	0.038
<i>Series 2</i>			
Distance from Community + Mean Price + Distance from Community:Mean Price	5	0.00	0.106
Body Size + Distance from Community + Mean Price + Distance from Community:Mean Price	6	1.11	0.061
Success Rate + Distance from Community + Mean Price + Distance from Community:Mean Price	6	1.96	0.040
Group Size + Distance from Community + Mean Price + Distance from Community:Mean Price	6	1.97	0.039
Hunt Duration + Distance from Community + Mean Price + Distance from Community:Mean Price	6	1.99	0.039
Consumer/Producer Ratio + Distance from Community + Mean Price + Distance from Community:Mean Price	6	1.99	0.039
<i>Series 3</i>			
Distance from Community + Hunt Duration + Group Size + Mean Price + Success Rate + Distance from Community:Mean Price + Mean Price:Success Rate	10	0.00	0.068
Hunt Duration + Group Size + Mean Price + Success Rate + Mean Price:Success Rate	8	0.00	0.068

Distance from Community + Hunt Duration + Mean Price + Success Rate + Distance from Community:Mean Price + Mean Price:Success Rate	9	1.10	0.039
Distance from Community + Hunt Duration + Group Size + Mean Price + Distance from Community:Mean Price	8	1.29	0.036
Hunt Duration + Group Size + Mean Price	6	1.34	0.035
Distance from Community + Hunt Duration + Group Size + Mean Price + Success Rate + Distance from Community:Mean Price	9	1.39	0.034
Hunt Duration + Group Size + Mean Price + Success Rate	7	1.46	0.033
Hunt Duration + Group Size + Mean Price + Success Rate + Distance from Community + Mean Price:Success Rate	9	1.55	0.031
Group Size + Mean Price + Success Rate + Distance from Community + Distance from Community:Mean Price + Mean Price:Success Rate	9	1.66	0.030
Group Size + Mean Price + Success Rate + Mean Price:Success Rate	7	1.78	0.028
Hunt Duration + Mean Price + Success Rate + Mean Price:Success Rate	7	1.80	0.027
Consumer/Producer Ratio + Distance from Community + Hunt Duration + Group Size + Mean Price + Success Rate + Distance from Community:Mean Price + Mean Price:Success Rate	11	1.84	0.027
Consumer/Producer Ratio + Hunt Duration + Group Size + Mean Price + Success Rate + Mean Price:Success Rate	9	1.85	0.027
<i>Series 4</i>			
Distance from Community + Success Rate	4	0.00	0.539
Distance from Community + Success Rate + Species Group	7	0.33	0.458

Figure S1. Map of interview results and reported encounter locations with 17 hunters in Sucusari, Peru.



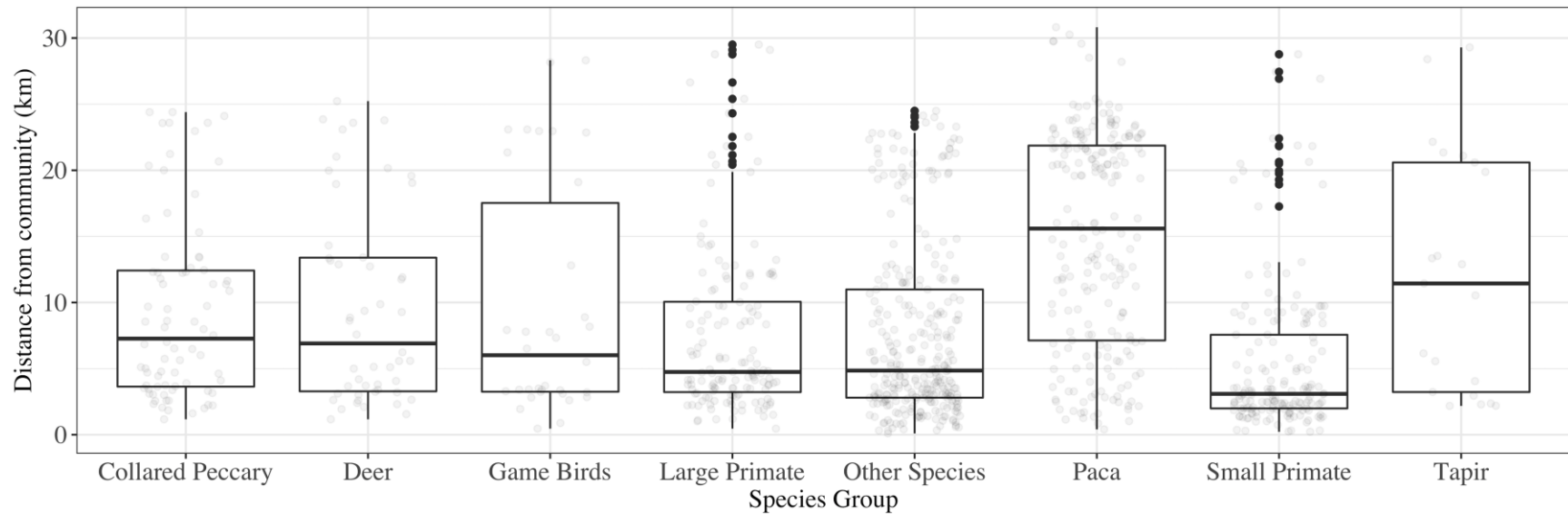


Figure S2. Boxplot of distribution of encounter locations' distance from the community by species group. Raw data shown by grey points.



## References

- Bardales García, J., Acosta, N. B., & Olivares, L. V. (2004). Técnicas de preservación y factor de conversión de fauna silvestre en la región Loreto, Perú. *Resúmenes VI Congreso Internacional Sobre Manejo de Fauna Silvestre En La Amazonía y Latinoamérica. Universidad Nacional de La Amazonía Peruana*, 427–433.
- Begazo, A. J., & Bodmer, R. E. (1998). Use and conservation of Cracidae (Aves: Galliformes) in the Peruvian Amazon. *Oryx*, 32(4), 301–309. <https://doi.org/10.1046/j.1365-3008.1998.d01-60.x>
- Bodmer, R. E., Pezo Lozano, E., & Fang, T. G. (2004). *Economic analysis of wildlife use in the Peruvian Amazon*.
- Dunning, J. B. (2007). *CRC Handbook of Avian Body Masses*. CRC Press.
- Fragoso, J. M. (2004). A long-term study of white-lipped peccary (*Tayassu pecari*) population fluctuation in northern Amazonia. *People in Nature, Wildlife Conservation in South and Central America*, 286–296.
- Kattan, G. H., Muñoz, M. C., & Kikuchi, D. W. (2016). Population densities of curassows, guans, and chachalacas (Cracidae): Effects of body size, habitat, season, and hunting. *The Condor: Ornithological Applications*, 118(1), 24–32.
- Loyola, R. D., Oliveira, G. D., Diniz-Filho, J. A. F., & Lewinsohn, T. M. (2008). Conservation of Neotropical carnivores under different prioritization scenarios: Mapping species traits to minimize conservation conflicts. *Diversity and Distributions*, 14(6), 949–960. <https://doi.org/10.1111/j.1472-4642.2008.00508.x>

Robinson, J. G., & Redford, K. H. (1986). Body size, diet, and population density of Neotropical forest mammals. *The American Naturalist*, 128(5), 665–680.